The synergistic effect in coal/biomass blend briquettes combustion on elements behavior in bottom ash using ICP-OES

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This paper focuses on the study of the synergistic effect in coal/biomass blend briquettes combustion on behavior of Al, As, Ba, Cd, Co, Cr, Cu, Fe, Ga, K, Mn, Mo, Ni, P, Pb, Si, V, W, Zn, Zr and characterization of raw materials and bottom ashes. The manufacturing of coal/biomass briquettes although not commonly used is an attractive approach, as briquettes combustion is more technologically advantageous than the fluidized bed combustion. In the same time this technology is a way to render valuable materials of low calorific power and results in diminishing polluting emission. Raw materials and briquettes from different blends of pitcoal/sawdust were subjected to combustion in a 55 kW-boiler. The total content of elements after digestion in the $HNO_3 - HF$ mixture and the content in water leachate at a solid/liquid ratio of 1:2 were determined both in raw materials and bottom ash by ICP-OES. The total content of elements was higher in pitcoal than in sawdust. The synergistic effect depends both on coal/biomass ratio in blend and element nature. The water leachable fraction of elements from ash decreased along with the increase of sawdust weight excepting macronutrients (K, P) and Si.

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1. Introduction

Coal/biomass blends are increasingly important in meeting targets for renewable energy utilization [1-5]. In connection with the EU regulations stipulating the utilization of renewable sources for energy production, Romania will have to triple in the next years its energy production capacities from renewable sources. A promising way is the employment of by-products from wood processing industry as energy source. Although having lower calorific power than coal, biomass is considered an environmentally safe way of providing energy, especially for process heat and district-heating purposes. The use of biomass for industrial purposes is small but is a strategic resource in the effort to fulfill the Kyoto agreement to replace fossil fuels and to mitigate greenhouse gas emissions. The most widespread use of biomass in energy industry is co-combustion with coal by various technologies like cofiring, fluidized bed or in form of briquettes of coal and biomass [5, 6]. Although not commonly used the combustion of coal/biomass briquettes represents an attractive approach to provide energy for industrial and domestic purposes. The advantages are related to similar rate of combustion to that of coal, a uniform combustion, reduced particulate emission. Moreover the same coal combustion facilities can be used without major modifications [7]. The process of briquetting consists of coal and biomass

mixing together with a binding substance of residual origin (molasses, oily organic wastes, starch, etc.) than pressing [2, 8]. The ash resulted from coal combustion alone or in mixture with biomass should not be regarded only as waste, since it could be used in industrial and agricultural purposes [5]. Ashes contain heavy metals, which could be released by leaching when residues are disposed on municipal landfill and represent a matter of concern for the environment [9-11]. The concentration of the heavy metals in ash depends on the fuel nature, ash type, namely bottom or fly ash as well as technology of firing [7].

Co-combustion raises the question of combustion products, their influence on environment and the synergistic effects that may change the composition of burning emission and bottom ash. The biomass favors the release of hydrogen sulphide during the thermal treatment. This fact can be explained in terms of the hydrogen-donor character of the biomass [12]. The synergistic effect is also important in relation with both organics and heavy metals. Devolatilization characteristics of coal and biomass blends show synergistic effects regarding the behavior of aromatics and phenols during burning. The amplitude of effects depends on type of coals and biomass in the fuel blend and pyrolysis technique, respectively [13]. The organic compounds of interest in the study of co-combustion effects are: polycyclic aromatic hydrocarbons (PAH), alkyl PAH, phenols, aldehydes, ketones, oxygenated and nitrogenated polycyclic aromatic compounds (O-PAC, N-PAC), dioxins, polycyclic aromatic sulphur hydrocarbons (PASH), common

volatile organic compounds (VOC), inorganic gases (CO, CO₂, SO_x and NO_x) [14, 15]. The synergistic behavior of heavy metals in coal/biomass co-combustion processes is not studied.

Main objective of this study is finding a clean coal utilization combined with biomass (sawdust) as briquettes for energy generation. For this purpose the synergistic effect in pitcoal/sawdust blend briquettes combustion on several elements behavior (Al, As, Ba, Cd, Co, Cr, Cu, Fe, Ga, K, Mn, Mo, Ni, P, Pb, Si, V, W, Zn and Zr) was studied by their monitoring in raw materials and bottom ash using inductively coupled plasma optical emission spectrometry (ICP-OES). The influence of the combustion process was evaluated by determining the total element content and that extracted in water leachates of raw materials and bottom ash. The mass balance of elements allowed the assessment of the elements enrichment in bottom ash.

2. Experimental

2.1 Materials and briquettes manufacturing

The block diagram of the technological process of bio-briquettes manufacturing is presented in Fig. 1. The materials used were granulated pitcoal coming from the Jiului Valley (Roamania) and wastes from wood industry (sawdust of resinous, walnut, oak, etc.). The pitcoal powder and sawdust were dried at 65° C than mixed in different ratios (w/w) in the range of 10-75 % sawdust, molasses and lime as binding additives and the mixture was shaped into briquettes under a pressure of $10 - 30 \text{ kN/m}^2$. The presence of sawdust provided an improved mechanical strength and allowed a higher pressure to be applied in order to lessen the volume of briquettes.

2.2 Burning system

The combustion of pitcoal, sawdust of resinous, beech and oak and briquettes containing pitcoal and resinous sawdust was conducted in a 55 kW-boiler (PIFATI-SA) located within the Laboratory of Boilers and Combustion Installations of the Department of Classic Thermo-mechanic and Nuclear Equipment, University Politehnica Bucharest. Combustion was conducted under natural convection conditions with air in excess. The boiler volume is 0.25 m³ (750 mm lenghtx550 mm widthx600 mm height). It has a cast iron grid for combustion with a slight slope towards the supply door, [4]. The energy delivered by the boiler provides the warm-up of a space volume of 500 m³, the equivalent of an individual housing. The amount of the bottom ash decreased along with the increase of the sawdust fraction in the blend subjected to combustion and was in the range of 2.0 - 13.5 % of the combusted weight.



Fig. 1. Block diagram of the technological process of bio-briquettes manufacturing.

2.3 Instrumentation

The determinations were carried out using the ICP multichannel spectrometer SPECTRO CIROS^{CCD}

(Spectro Analytical Instruments Kleve, Germany). Details about operating conditions are given in Table 1. The wavelengths and the limits of detection for the ICP - OES are presented in Table 2.

| Equipment | SPECRO CIROS ^{CCD} |
|-------------------------------|--|
| Generator | Free – runing 27.12 MHz operated at 1400 W. |
| Plasma torch | Inductively coupled plasma, axial viewing; torch positioning (mm): $X = -3.9$; $Y = +3.6$; $Z = +$ 2.6. Argon flow rates: Outer gas 12 L min ⁻¹ Intermediate gas 0.6 L min ⁻¹ Nebulizer gas 1 L min ⁻¹ |
| Sample introduction system | 4 channel peristaltic pump, K2 cross – flow nebulizer, double pass Scott type spray chamber. Sample uptake rate: 2 mL min ⁻¹ Flushing time: 40 s Delay time: 20 s |
| Optics | 160 – 800 nm double grating Paschen – Runge multichannel spectrometer. Chamber filled with Ar. |
| Detector | 22 Charge Coupled Detectors (CCDs) |
| Data processing | Smart Analyzer Software. Background correction: linear and square two points models, best SNR strategy, integration time 45 s and 3 successive measurements for each parallel sample. |

Table 1. Operating parameters for the ICP – OES SPECTRO CIROS^{CCD}.

Table 2. Wavelengths and limits of detection (3σ criteria) in ICP - OES.

| Element | λ / nm | LOD/ ng ml ⁻¹ | Element | λ / nm | LOD/ ng ml ⁻¹ |
|---------|----------------|--------------------------|---------|----------------|--------------------------|
| Al | 396.152 | 4 | Mn | 257.610 | 0.4 |
| As | 189.042 | 6 | Мо | 379.825 | 10 |
| Ba | 455.404 | 0.2 | Ni | 341.476 | 7 |
| Cd | 214.438 | 1 | Р | 213.618 | 3.5 |
| Co | 238.892 | 1.5 | Pb | 220.351 | 35 |
| Cr | 283.563 | 14 | Si | 212.412 | 7 |
| Cu | 324.754 | 1.5 | V | 292.402 | 4 |
| Fe | 261.187 | 5 | W | 239.709 | 20 |
| Ga | 403.299 | 12 | Zn | 213.856 | 0.4 |
| K | 766.490 | 11 | Zr | 343.823 | 0.1 |

A mortar grinder Restch RM 100 and a sieve shaker Restch AS 200 (Haan, Germany), an overhead shaker REAX20/8 Heidolph (Kelheim, Germany), a Memmert UFE 500 oven (Schwabach, Germany) and a closed – vessel microwave system Berghof MWS-3+ with temperature control mode (Eningen, Germany) were used during preparation of the analytical samples.

2.4 Reagents and standard solutions

A stock solution of 1000 μ g mL⁻¹ of Al, Ba, Cd, Co, Cr, Cu, Fe, Ga, K, Mn, Ni, Pb, and Zn and another of 1000 μ g mL⁻¹ of As, Mo, Si, P, V, W and Zr purchased from Merck (Darmstadt, Germany) were used in this study for preparation of calibration standards. Standard solutions in the range of 0 – 5 μ g mL⁻¹ element were prepared by serial dilution of the stock standard solutions with 5 % HNO₃ (v/v) and high purity water (18.2 MΩ_. cm⁻¹) from a Milli Q system (Millipore, Milford, MA, USA). Ultrapure nitric acid 65 %, Suprapure hydrofluoric acid 40 % purchased from Merck (Darmstadt, Germany), and boric acid 99.9 % purchased from Carlo Erba Reagenti (Milano, Italy) were used for digestion of samples. Argon (5.0 quality) from Linde Gas SRL Cluj – Napoca, Romania was used as working gas for ICP generation.

2.5 Analytical sample preparation

The ash, coal, sawdust and minced briquettes samples were dried in an oven at 105 ± 5^{0} C for 2 h. After cooling the samples were sieved through a 4 mm sieve and the remaining fraction was grinded down until more than 95 % of the sample was sieved through. Three replicates of 175 g of sample prepared in such way were leached with high purity water at solid-to-liquid ratio of 1:2. The leaching was performed in a Heidolph shaker at 16 rpm, for 24 ± 0.5 h at 20 ± 2^{0} C (SR EN 12457/1:2003). The total content of elements was determined based on 3 replicates of sample (< 250 µm) subjected to the acid digestion using the reagents volumes mentioned in Table 3.

Elements were determined in solutions by ICP-OES.

| Sample | Mass (g) | 1 st Sta | 2 nd Stage | |
|------------------------|----------|---|-----------------------|-----------------------------------|
| | | $V_{\rm HNO3}65\%({\rm mL})$ $V_{\rm HF}40\%({\rm mL})$ | | V _{H3BO3} saturated (mL) |
| Pitcoal | 0.2000 | 9 | 4 | 30 |
| Sawdust and briquettes | 0.3000 | 10 | 3 | 30 |
| Ash | 0.4000 | 9 | 4 | 30 |

Table 3. Scheme for sample preparation by microwave digestion for total element content.

The digestion program in closed PTFE containers of the microwave system is presented in Table 4.

| Ash | | | | | | | | |
|------------------------------|---------------------|-----------------------|-------------|---------------|-----|-----|--|--|
| | 2 nd Sta | 2 nd Stage | | | | | | |
| | | | | Steps | | | | |
| Parameter | 1 | 2 | 3 | 4 | 1 | 2 | | |
| Temperature / ⁰ C | 180 | 230 | 230 | 100 | 200 | 100 | | |
| Ramp time / min | 2 | 2 | 1 | 1 | 5 | 1 | | |
| Hold time / min | 7 | 19 | 19 | 5 | 15 | 15 | | |
| Power / %* | 90 | 90 | 90 | 10 | 90 | 10 | | |
| | | C | oal, sawdus | t, briquettes | | | | |
| Temperature / ⁰ C | 150 | 180 | 180 | 100 | 200 | 100 | | |
| Ramp time / min | 2 | 2 | 1 | 1 | 5 | 1 | | |
| Hold time / min | 7 | 19 | 19 | 5 | 15 | 15 | | |
| Power / %* | 90 | 90 | 90 | 10 | 90 | 10 | | |

| Table 4. Operating | conditions | for the microwave | digestion system. |
|--------------------|------------|-------------------|-------------------|
| | | | |

*100 % power corresponds to 1450 W

3. Results and discussion

3.1 Characterization of raw materials and bottom ashes

The range and average values of major, minor and trace elements in pitcoal, sawdust and bottom ash are summarized in Tables 4-6. The contents of As, Ba, Cd, Co, Cr, Ga, Ni, Pb and W in pitcoal were over the average values for most world coals, while those of Cu, Mo, P, V, Zn and Zr were similar [16]. Among major elements, Al and K occurred in higher concentrations than reported for

feed coals used in a power plant in Turkey but Fe fall in the same range [17].

Higher concentrations of both major and trace elements were found in pitcoal than in sawdust used as raw materials in the briquettes manufacturing. The contents of heavy metals found in bottom ash are comparable to those reported in the literature [7]. The major elements (Al, Fe, Si) and some traces (As, Co, Cu, Ni, V, W and Zr) were found in higher concentration in ash resulted from pitcoal, while in sawdust ash elements considered as macronutrients (K, P) as well as Ba, Cd, Cr, Ga, Mn, Mo, Pb and Zn were present in higher concentrations.



Fig. 2. Comparative extraction degrees in water of several elements from pitcoal, pitcoal ash, sawdust and sawdust ash at a solid/liquid ratio of 1:2.

| | | Pit | coal | | Pitcoal ash | | | |
|---------|---------|--------|------|------|-------------|--------|-------|-------|
| Element | Average | Median | Min | Max | Average | Median | Min | Max |
| % Al | 7.4 | 7.5 | 6.9 | 8.0 | 7.3 | 7.5 | 6.6 | 7.8 |
| % Fe | 2.3 | 2.3 | 2.2 | 2.4 | 2.7 | 2.7 | 2.4 | 3.0 |
| % K | 1.7 | 1.8 | 1.5 | 1.9 | 1.6 | 1.9 | 1.0 | 2.1 |
| % Si | 14.1 | 13.9 | 12.6 | 16.0 | 8.3 | 9.3 | 5.9 | 9.6 |
| As | 328 | 324 | 309 | 355 | 314 | 323 | 293 | 324 |
| Ва | 2000 | 1690 | 1020 | 3585 | 2106 | 2089 | 2040 | 2190 |
| Cd | 19 | 18 | 18 | 22 | 37 | 36 | 36 | 39 |
| Со | 240 | 237 | 233 | 252 | 304 | 308 | 273 | 329 |
| Cr | 118 | 117 | 105 | 132 | 177 | 173 | 162 | 195 |
| Cu | 50 | 51 | 44 | 53 | 411 | 388 | 311 | 534 |
| Ga | 272 | 269 | 242 | 307 | 262 | 269 | 232 | 283 |
| Mn | 281 | 277 | 248 | 319 | 224 | 228 | 201 | 242 |
| Мо | 32 | 33 | 25 | 37 | 35 | 36 | 28 | 41 |
| Ni | 110 | 108 | 93 | 129 | 225 | 228 | 204 | 242 |
| Р | 476 | 478 | 426 | 521 | 477 | 485 | 438 | 507 |
| Pb | 416 | 402 | 383 | 477 | 1169 | 1196 | 1100 | 1210 |
| V | 87 | 89 | 76 | 96 | 109 | 107 | 96 | 123 |
| Zn | 126 | 122 | 98 | 163 | 12910 | 12850 | 12490 | 13390 |
| Zr | 92 | 89 | 77 | 115 | 94 | 94 | 86 | 102 |
| W | 167 | 167 | 158 | 175 | 202 | 209 | 184 | 213 |

Table 5. Total content of elements in pitcoal and pitcoal ash samples after acidic digestion (mg/Kg unless indicated otherwise).

Table 6. Total content of elements in sawdust and sawdust ash samples after acidic digestion (mg/Kg unless indicated otherwise).

| | | Saw | 'dust | | Sawdust ash | | | | |
|---------|---------|--------|-------|------|-------------|--------|-------|-------|--|
| Element | Average | Median | Min | Max | Average | Median | Min | Max | |
| % Al | 0.25 | 0.11 | 0.07 | 8.8 | 2.0 | 5.1 | 2.4 | 5.5 | |
| % Fe | 0.11 | 0.06 | 0.04 | 0.28 | 1.7 | 1.6 | 1.4 | 2.2 | |
| % K | 0.28 | 0.14 | 0.13 | 0.80 | 3.9 | 4.4 | 2.1 | 5.2 | |
| % Si | 5.6 | 2.2 | 1.8 | 19.8 | 7.2 | 6.8 | 4.0 | 10.7 | |
| As | 161 | 150 | 145 | 211 | 157 | 152 | 114 | 203 | |
| Ba | 707 | 403 | 378 | 1790 | 6900 | 6900 | 5670 | 8135 | |
| Cd | 6 | 5 | 4 | 10 | 41 | 46 | 22 | 55 | |
| Со | 3 | 3 | 1 | 5 | 224 | 233 | 200 | 240 | |
| Cr | 5 | 6 | 1 | 7 | 352 | 403 | 185 | 470 | |
| Cu | 3 | 3 | 2 | 5 | 170 | 178 | 127 | 205 | |
| Ga | 22 | 21 | 3 | 48 | 1290 | 1290 | 782 | 1805 | |
| Mn | 41 | 15 | 1 | 150 | 1080 | 1030 | 610 | 1605 | |
| Мо | 12 | 11 | 11 | 19 | 42 | 39 | 28 | 57 | |
| Ni | 11 | 7 | 3 | 22 | 39 | 45 | 3 | 114 | |
| Р | 98 | 65 | 51 | 183 | 2380 | 1900 | 1850 | 3400 | |
| Pb | 157 | 152 | 128 | 207 | 4210 | 4090 | 3390 | 5155 | |
| V | 3 | 4 | 1 | 5 | 52 | 48 | 44 | 62 | |
| Zn | 41 | 32 | 15 | 80 | 21400 | 19870 | 19495 | 24850 | |
| Zr | 56 | 19 | 14 | 212 | 71 | 67 | 50 | 95 | |
| W | 18 | 6 | 6 | 64 | 137 | 156 | 86 | 170 | |

The supraunitary ratio of elements concentrations in ash/raw material, higher for sawdust than pitcoal, shows a grater enrichment of elements in sawdust ash.

The analysis of raw materials and corresponding bottom ash leachates revealed an increase of water mobility following combustion for most of elements. Element release was higher from pitcoal ash compared to sawdust ash excepting K, P and Si (Fig. 2). In relation with OM 95/2005 regarding waste classification, bottom ashes comply with requirements for non-dangerous waste and could be used as amendments on agricultural land.

| | | Briq | uettes | | | Briquettes ash | | | | |
|---------|---------|--------|--------|-----|---------|----------------|------|-------|--|--|
| Element | Average | Median | Min | Max | Average | Median | Min | Max | | |
| % Al | 6.7 | 7.0 | 6.0 | 7.2 | 6.3 | 6.3 | 5.0 | 7.4 | | |
| % Fe | 1.7 | 1.8 | 1.4 | 1.9 | 1.7 | 1.5 | 1.4 | 2.4 | | |
| % K | 1.4 | 1.3 | 1.2 | 1.6 | 1.9 | 2.2 | 1.0 | 2.6 | | |
| % Si | 8.5 | 8.5 | 7.7 | 9.4 | 11.8 | 11.6 | 11.1 | 12.8 | | |
| As | 224 | 237 | 187 | 246 | 200 | 195 | 167 | 233 | | |
| Ba | 855 | 840 | 820 | 905 | 2890 | 2730 | 1960 | 4150 | | |
| Cd | 12 | 17 | 4 | 17 | 26 | 29 | 20 | 30 | | |
| Co | 170 | 200 | 114 | 201 | 200 | 175 | 170 | 283 | | |
| Cr | 83 | 94 | 58 | 98 | 114 | 62 | 59 | 275 | | |
| Cu | 45 | 51 | 30 | 55 | 91 | 93 | 80 | 99 | | |
| Ga | 147 | 175 | 80 | 186 | 1950 | 2380 | 367 | 2670 | | |
| Mn | 146 | 160 | 115 | 166 | 1550 | 1880 | 355 | 2080 | | |
| Мо | 24 | 27 | 13 | 31 | 28 | 17 | 24 | 35 | | |
| Ni | 93 | 134 | 3 | 141 | 104 | 78 | 75 | 184 | | |
| Р | 265 | 272 | 236 | 290 | 2140 | 2390 | 1205 | 2575 | | |
| Pb | 263 | 390 | 9 | 395 | 960 | 754 | 715 | 1630 | | |
| V | 66 | 69 | 57 | 73 | 61 | 43 | 41 | 116 | | |
| Zn | 247 | 305 | 93 | 345 | 6960 | 4750 | 4135 | 14200 | | |
| Zr | 60 | 61 | 55 | 62 | 63 | 53 | 51 | 97 | | |
| W | 105 | 112 | 78 | 126 | 190 | 184 | 183 | 200 | | |

Table 7. Total content of elements in briquettes and briquettes ash after acidic digestion (mg/Kg unless indicated otherwise).

3.2 Synergistic effect in pitcoal-sawdust blend combustion

In order to study the fate of elements during combustion of pitcoal/sawdust blends, their concentrations were determined in the bottom ashes resulted from briquettes containing 0; 10; 25; 50; 75 and 100% sawdust. Results are plotted in Fig. 3. According to [18], when the content of an individual component in the combusted product is linearly proportional to the blending ratio there is no synergistic effect between coal and biomass over the selected range of experimental conditions

As shown in Fig. 3, the synergistic effect on elements behavior during combustion of pitcoal/sawdust blend was complex depending on the composition of the mixture subjected to burn and element type. For a group of nine elements (Al, As, Cd, Co, Cr, Fe, K, Ni and V) the synergistic effect was the highest. For this group of elements it was observed a maximum content in the bottom ash for briquettes containing 25-50% sawdust. For the rest of elements (Ba, Cu, Ga, Mn, Mo, P, Pb, Si, W, Zn and Zr) the synergistic effect was less evident since their concentration in bottom ash increased almost continuously with the increase of the sawdust weight in the blend. Generally, the synergistic effect decreased with the increase of sawdust fraction in the briquettes. The range of recovery degrees of elements in bottom ash calculated based on amounts corresponding to parent material and ash is presented in Table 8.Elements can be divided into two groups considering their recovery degrees in bottom ash. To the first group belong elements of high recovery degree (Fe, Ba, Cd, Co, Cr, Cu, Ga, K, P, Pb, V, Zn), generally retained on Fe compounds. The second group comprises elements with lower recovery degree bound

mainly to Al and Si (Al, Si, As, Mo, Mn, Ni, W, Zr) suggesting that high amounts of them are volatilized during combustion and evacuated along with the gaseous products. The development of volatile compounds of As during combustion is responsible for its low recovery degree in bottom ash. Although considered as refractory elements Al, Mo, W, Si and Zr sublimate probably during combustion in a significant extent. From the point of view of elements enrichment in bottom ash, the manufacturing of briquettes containing 75% pitcoal and 25% resinous sawdust seems advantageous. The recovery degrees of elements in ash from the combustion of a briquettes charge with this composition are shown in Fig. 4.

According to Fig. 4, recovery degrees over 20% were found for Ba, Cd, Cu, Ga, K, P, Pb and Zn in bottom ash of briquettes and were within the range of those obtained for the bottom ashes of different mixtures of parent materials.

The synergistic effect on elements availability from bottom ash resulted from briquettes combustion was investigated based on the extraction degrees in water related to total contents. The extraction degrees for several elements from bottom ash in water for a solid/liquid ratio 1:2 are presented in Fig. 5.

The synergistic effect is observable through the nonlinear change in the extraction degree of elements from ash in water along with the composition of the blend subjected to burn. The leachability of elements seen as essential plant nutrients (K, P) increased up to 8 and 0.8% respectively as the biomass fraction in the blend increased. The reason is that these elements are absorbed from soil by plants as water soluble species and are found further in ash also in soluble form. The extraction degree of Si from sawdust bottom ash in water was appreciable and could seem anomalous at first sight, however it agrees with literature data. The soil water contains Si as silicic acid in concentration on the order of major plant nutrients and is readily absorbed together with macronutrients. Consequently Si concentration in plants is appreciable ranging from 1 to 10 % or even higher [19]. The water

availability of other elements from pitcoal ash was (%): 10 (Ga, Mn), 1 (Zn), 0.5 (Co), 0.2 (Cd), 0.1 (Cu) and decreased as the sawdust in the mixture increased. The same tendency was found for elements with possible toxic effect (As, Cd, Pb) but with water extraction degrees below 0.01 %.



Fig. 3. Total element contents in bottom ash from combustion of briquettes of different compositions.

| Element | Average | Median | Minimum | Maximum | Element | Average | Median | Minimum | Maximum |
|---------|---------|--------|---------|---------|---------|---------|--------|---------|---------|
| Al | 31.9 | 23.3 | 14.9 | 62.4 | Cd | 55.7 | 65.8 | 17.9 | 93.1 |
| Κ | 50.7 | 58.7 | 5.2 | 93.1 | Ва | 42.9 | 28.8 | 16.0 | 86.4 |
| Cr | 47.8 | 43.5 | 17.8 | 85.5 | Pb | 71.1 | 74.8 | 37.3 | 91.6 |
| Mn | 31.0 | 27.7 | 12.1 | 62.4 | Si | 23.6 | 23.5 | 11.2 | 38.7 |
| Fe | 48.5 | 38.8 | 18.2 | 92.7 | Р | 50.2 | 41.5 | 15.2 | 94.1 |
| Со | 42.9 | 38.1 | 15.3 | 72.4 | V | 48.7 | 48.6 | 18.9 | 80.4 |
| Ni | 33.0 | 31.2 | 18.5 | 48.1 | As | 15.5 | 15.2 | 2.9 | 24.1 |
| Cu | 67.4 | 75.9 | 20.2 | 95.5 | Zr | 18.5 | 15.4 | 6.8 | 27.9 |
| Zn | 40.9 | 39.5 | 19.8 | 71.3 | Мо | 24.2 | 16.4 | 13.8 | 42.3 |
| Ga | 42.3 | 44.9 | 13.7 | 91.8 | W | 36.2 | 34.5 | 18.4 | 54.8 |

Table 8. Range of recovery degrees (%) of elements in bottom ash from combustion of briquettes.

The element fractions extracted in water from bottom ash of briquettes containing 75% pitcoal and 25% resinous sawdust are plotted in Fig. 6. The extraction degrees in water of elements from bottom ash of briquettes containing 75% pitcoal and 25% resinous sawdust (Fig.6) were similar to those obtained in the analysis of the ashes presented above (Fig.3). The extraction degrees of toxic elements such as As, Cd and Pb were very low as a result of their presence as water insoluble species in ash.



Fig. 4. Recovery degrees of elements in bottom ash from combustion of briquettes containing 75 % picoal and 25 % resinous sawdust.



Fig. 5. Extraction degrees of elements from bottom ash by water leaching for a solid/liquid ratio 1:2



Fig. 6. Extraction degrees of elements by water leaching of ash resulted from briquettes containing 75% pitcoal and 25% resinous sawdust for a solid/liquid ratio 1:2

4. Conclusions

The study demonstrates that the synergistic effect in coal/biomass briquettes combustion on the total content and water leachable fraction of elements in bottom ash depends on the combusted blend and element type. The synergistic effect was higher for Al, As, Cd, Co, Cr, Fe, K, Ni, V and lower for Ba, Cu, Ga, Mn, Mo, P, Pb, Si, W, Zn and Zr. Characterization of parent materials emphasized a total content of elements in pitcoal higher than in sawdust. The enrichment of elements was higher in sawdust bottom ash compared to pitcoal. The mass balance revealed recovery degrees in bottom ash up to 70-95 % for Ba, Cd, Co, Cr, Cu, Fe, Ga, K, Pb, V and Zn and lower, up to 30-50% for As, Al, Mn, Mo, Ni, Si, W, Zr. The leachability study emphasized higher extraction degrees for Cd, Co, Cu, Ga, Mn and Zn in pitcoal bottom ash and their decrease along with the increase of the biomass fraction in the combusted blend. The extraction degrees of macronutrients (K, P) and Si were maximum for sawdust ash. For the rest of elements under study, of which some toxic (As, Cd, Pb), extractions were below 0.01%, corresponding to concentrations below those stipulated by the legislation covering the transport, storage and disposal of wastes. The manufacturing of coal/biomass blend briquettes provides a promising approach to render valuable materials of low calorific power ensuring in the same time environment protection against pollution with toxic elements as easily water leachable species.

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